

Environmental Behavior of the Herbicide Acetochlor in Soil

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Acetochlor [2-chloro-N-(ethoxymethyl)-N-(2-ethyl-6-methylphenyl) acetamide] is one of the most widely applied herbicides for controlling competing grasses and some broadleaf weeds in corn field (Capel, 1995), and has been classified as a B-2 carcinogen by the USEPA (USEPA, 1994). About 10 million kg of acetochlor (active ingredient) has been applied every year since 1977 in China (Wang, 1999). Acetochlor was one of the most widely used agricultural herbicides in China.

The environmental aspects of acetochlor use are of current concern because of its presence in surface and ground waters in most agricultural regions (Karlkhoff, 1998; Clark, 1999). This is particularly the case with acetochlor, where a considerable portion of the applied herbicide may be lost from the target, mostly due to processes such as adsorption, degradation, runoff and leaching. The investigation of environmental behavior and fate of acetochlor in soils is of great significance for raising utilization efficiency and reducing groundwater contamination of this herbicide.

Through field and laboratory experiments, the processes of adsorption, biodegradation, hydrolysis, and dissipation dynamics of acetochlor in soils were studied. We also investigated influences of humic acid (HA) and anionic surfactant sodium dodecylbenzene sulfonate (SDBS) on the behavior of acetochlor in soils. The studied field was located at the corn field of Northwestern Beijing, where acetochlor has been applied for about 10 years.

MATERIALS AND METHODS

Soil samples were taken from top 10 cm layer at the agricultural field of Northwestern Beijing that did not receive previous application of the herbicide. The bulk soil was slightly air-dried, thoroughly mixed, and passed through 0.15 mm sieves. Analyzed soil properties include pH, organic carbon (OC), cation exchange capacity (CEC) and particle-size distribution (Table 1).

Acetochlor in soil and water was analyzed by LC-6A High Performance Liquid Chromatography (HPLC) with Shimadzu Spectrophotometric Detector (SPD-3). The detection wavelength was 210 nm. The mobile phase was methanol-water (80/20, v/v) with a flow rate of 1.0 ml/min.

The procedures of extracting acetochlor from soil samples were: 20 g of soil sample and 40 ml acetone solution of 80% were placed into 500 ml of centrifuge tube, and extracted on ultrasonic oscillator for 10 min. The suspended solution in centrifuge tube was separated with 3000 rpm of centrifuge. The above extraction procedure was repeated 3 times and three extracts were combined. The combined extract was concentrated on rotatory evaporator to remove most of the acetone. The concentrated residue was transferred into a 100 ml separated

Table 1. Physical and chemical properties of the soil

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PH	OC (%)	CEC (mmol/100g)	silt (%)	sandy (%)	clay (%)
8.4	1.5	84.0	59	38	3

funnel with 40 ml water and 3 g NaCl, and extracted two times with 30 ml petroleum ether. The two extracts were combined and concentrated near dry on rotatory evaporator, and then totally dried under a stream of N_2 for HPCL analysis.

The water samples were extracted into petroleum ether and concentrated to about 3 ml under reduced pressure. The concentrated extract was directly dried under a gentle stream of nitrogen gas and the residue was dissolved in 1 ml of methanol, and then analyzes by HPLC.

Standard solution (1.42 ml 352 mg.l⁻¹) of acetochlor was taken into test tubes and were diluted to 25 ml using pH4, pH7 and pH10 distilled waters containing 1% NaN₃ as a biocide respectively. All test tubes were screw caped, and placed into an incubator at 25 ± 1 °C. $20\,\mu$ l sample was taken from test tubes for acetochlor measurement at regular intervals.

Twenty g soil sample was placed into a 100 ml-Erlemeyer, in which acetochlor was added and mixed homogenously to obtain test sample that contained 10 mg.kg⁻¹ acetochlor. Samples were moistured with asepsis distilled water and kept suitable moisture, and the top of the Erlemeyer was sealed with asepsis plastic membrane. The samples were incubated at 35±1 °C. Samples were taken out at regular intervals and extracted for acetochlor analysis. Another set of samples was kept under asepsis condition as control experiments.

Adsorption distribution coefficients of acetochlor for soil samples were determined in duplicate using the batch equilibration method. For each determination, certain amount of soil samples were weighed into 50-mL glass centrifuge tubes and 10 ml of aqueous acetochlor solution of known concentrations (20, 40, 60, 80, and 100 mg Γ^1) was added to each tube. The acetochlor aqueous solutions contained 1% NaN $_3$ as a biocide. Tubes were sealed with Teflon-lined screw caps and mechanically shaken for 21 hr at 25 °C. After shaking, tubes were centrifuged for 15 min at 2500rpm. 5-mL aliquot of the supernatant was removed, filtered through 0.45 μm filter membrane and placed in a glass tube for HPLC analysis of acetochlor

RESULTS AND DISCUSSION

The hydrolysis kinetic process of acetochlor can be described by the first-order reaction law and its rates in distilled and river waters were almost the same. H⁺ can accelerate acetochlor hydrolysis. The hydrolysis rate constant in pH4 water was greater than in pH7 and pH10 waters. The hydrolysis half-live of acetochlor in pH4, pH7 and pH10 river waters were 1386 d, 2310 d and 2310 d, respectively.

The biodegradation behavior of a pesticide in soil is important in understanding its fate. The biodegradation processes of acetochlor in soils without and with HA and SDBS were measured during 35 days laboratory incubation. The biodegradation half-life in soils was 4.6 d. Addition of HA and SDBS decreased acetochlor biodegradation. The half-life was increased 17% and 33% by addition of HA and SDBS, respectively (Fig.1).

Acetochlor may form acetochlor-HA complex with HA. The complex has a much greater molecular and is easier to be adsorbed onto soil particles, and thus more difficult to be degraded by microorganism than under condition without HA. The biodegradation products of acetochlor in soil are very complex, but two compounds can be identified by GC/MASS, i.e. hydroxyacetochlor [2-hydroxy- N-(ethoxymethyl)-6'-ethylacet-o-toluidide] and

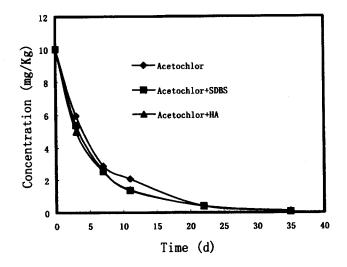


Figure 1. Biodegradation process of acetochlor in soil (Rate constants of biodegradation: Acetochlor only: 0.15 d⁻¹; Acetochlor + SDBS: 0.11 d⁻¹; Acetochlor + HA: 0.13 d⁻¹)

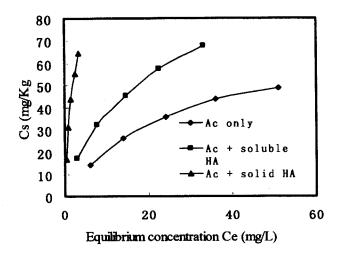


Figure 2. Adsorption isotherm of acetochlor to soil under soluble and solid HA (Ac: acetochlor)

2-methyl-6-ethylaniline.

In figure 2 and 3, Cs is the concentration of acetochlor adsorbed on soil ($mg.kg^{-1}$); Ce is the equilibrium concentration of acetochlor in solution ($mg.l^{-1}$). The calculated results were shown in table 2. The mean value of adsorption coefficient K_f was 4.339 L.kg⁻¹. Both solid and soluble HA increased adsorption capacity of acetochlor, but the former was greater than

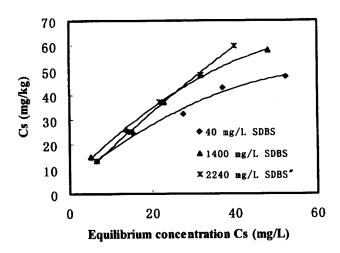


Figure 3. Adsorption isotherm of acetochlor to soil under SDBS

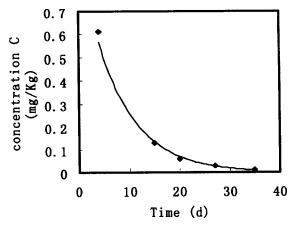


Figure 4. Dissipation process of acetochlor in surface 10 cm layer soil in corn field C=0.9603e^{-0.1313t}, R²=0.9964)

the latter. Solid HA had great influence on acetochlor adsorption because solid HA in soil may be a sorbent itself, which played an important part in the adsorption of acetochlor.

The acetochlor could be solubilized at low SDBS concentration. The adsorption of acetochlor to soil could be increased at higher SDBS concentration since the soil surface might be completely covered by a monolayer of surfactant and the additional surfactant monomer might be adsorbed on the monolayer of hydrophobic heads. This would result in a bilayer of surfactant on the soil surface with surfactant hydrophobic tails sticking out in the aqueous phase, which would considerably enhance acetochlor adsorption to soil.

Dissipation of acetochlor in field soils is a result of combined action of biodegradation, photolysis, volatilization, leaching, runoff, and plant uptake. But the main processes are microbial degradation and leaching. Photolysis is not a main process because it can only

occur in top 0.4 mm soils. The measured data of acetochlor concentration in soil samples, which were taken from the top 10 cm layer at corn field, were fitted to the first-order equation. The dissipation rate constant of acetochlor in surface layer soil was 0.13 d⁻¹ (Fig.4). Data in Fig.4 were average values of two year measurements

Table 2. Freundlich constants describing adsorption of acetochlor to soils with some substances.

Substances	Kf	1/n	R^2
Ac only	4.339	0.636	0.992
Ac + soluble HA	4.405	0.651	0.999
Ac + solid HA	9.572	0.572	0.996
Ac + 40 mg/L SDBS	4.303	0.62	0.989
Ac + 1400 mg/L SDBS	5.353	0.621	0.996
Ac + 2240 mg/L SDBS	5.489	0.789	0.995

Ac: Acetochlor

By comparing the dissipation rate (rate constant=0.13d⁻¹) and biodegradation rate (rate constant=0.15d⁻¹), we found that the biodegradation rate under experimental conditions was greater than that under field conditions. Data from laboratory biodegradation measurement need to be calibrated before being used in field model calculation.

Environmental behavior of acetochlor in soils depends on intrinsic chemical properties and extrinsic environmental factors, such as soil properties and climatic conditions. Intrinsic acetochlor properties, such as vapor pressure, water solubility, biological activity and resistance to chemical changes, indicate the tendencies of herbicide fate, while extrinsic factors, such as soil microbial activity, soil organic carbon content, soil temperature, solar radiation and rainfall, and co-occurred other substances (e.g. HA and surfactants) can greatly modify the fate of acetochlor.

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